THE IMPACT OF BICYCLE LANE CHARACTERISTICS ON BICYCLISTS' EXPOSURE TO TRAFFIC-RELATED PARTICULATE MATTER

Christine M. Kendrick [1],*, Adam Moore[2], Ashley Haire[2], Alexander Bigazzi[2], Miguel Figliozzi[2], Christopher M. Monsere[2], Linda George[1]

*Corresponding Author

[1]Environmental Science and Management
Portland State University
P.O. Box 751
Portland, OR
Portland, OR 97201
Email: kendricc@pdx.edu
georgeL@pdx.edu
Phone: 503-725-3861

[2]Department of Civil and Environmental Engineering
Portland State University
P.O. Box 751
Portland, OR 97201-0751
Email: adam.moore@pdx.edu
haire@pdx.edu
abigazzi@pdx.edu
figliozzi@pdx.edu
monsere@pdx.edu

Submitted to the 90th Annual Meeting of the Transportation Research Board
January 23-27, 2011

Submitted July 2010, Revised November 15, 2010

6,530 Words [4,530 + 7 figures x 250+ 1 table x 250]

ABSTRACT

Bicycling as a mode of transportation is increasingly seen as a healthy alternative to motorized transportation modes. However, in congested urban areas the health benefits of bicycling can be diminished by the negative health effects associated with inhalation of particulate matter. Particles of small size (ultrafine particles <0.1µm) are the most harmful even during short-duration exposures. Since vehicular exhaust is the major source of ultrafine particles, this research studies impacts of traffic levels and bicycle lane characteristics on bicyclists' exposure. Ultrafine particle exposure concentrations are compared in two settings: (a) a traditional bicycle lane adjacent to the vehicular traffic lanes and (b) a cycle track design with a parking lane separating bicyclists from vehicular traffic lanes. Traffic measurements were made alongside air quality measurements. It was observed that the cycle track design mitigates ultrafine particle exposure concentrations for cyclists. Results show statistically significant differences in term of exposure levels for the two bike facilities as well as correlations between traffic levels and exposure level differences. Results also suggest that ultrafine particle levels and spatial distribution may be sensitive to proximity to signalized intersections. Findings of this research indicate that in high traffic areas bicycle facility design has the potential to lower bicyclists' air pollution exposure levels.

INTRODUCTION

Bicycling as a mode of transportation is an increasingly attractive mode due to livability initiatives geared towards reducing traffic congestion and air pollution, attempts to increase physical exercise levels, and greenhouse gas concerns. As a result there has been a growing interest to increase municipal investments in bicycle infrastructure. Due to accessibility needs of commuters and cost constraints, most cycling facilities are located within the existing right-of-way of urban roadways. Cyclists in these facilities face a number of adverse effects brought on by their proximity to automobile traffic, including vulnerability to conflicts with motor vehicles and air quality concerns from tailpipe emissions.

Vehicular exhaust is the source of a multitude of air contaminants, including particulate matter (PM). Particulate matter of concern ranges in size from the largest, PM_{10} (diameter<10µm) and $PM_{2.5}$ (diameter<2.5µm), to microscopic ultrafine particles (UFP). Ultrafine particles have diameters smaller than 0.1µm. The majority of ultrafine particles present in an urban environment are the result of traffic emissions *(1-3)*.

Particle number concentrations, which are dominated by ultrafine particles, have been shown to be significantly higher next to a road *(4,5)*. Elevated levels of ultrafine particles are of a concern to bicycle commuters due to the associated health effects and increased respiration and absorption as compared to other road users *(6-9)*. For a given mass concentration (µg/m^3), ultrafine particles have 10^2 to 10^3 times higher surface area than fine particles with diameters in the 0.1-2.5 µm range and about 10^5 times more than coarse particles (2.5µm -10µm) *(10)*. This higher surface area can increase the potential for ultrafine particles to carry toxins into the human body. The small size allows for the deepest deposition of particles into the alveolar region of the lungs, pulmonary interstitial spaces, and possible passage into the circulatory system, and it has been shown that these particles accumulate over time in organ tissues *(11)*. The deep deposition of these small particles in high numbers can provoke inflammation which is linked to increased or exacerbated asthma and oxidative stress which is involved in cardiovascular and pulmonary disease. The presence of a high number of particles in the alveolus has been shown to be more critical to adverse effects and indicative of potential health impacts than total particle mass concentrations *(12-14)*. The human pulmonary and cardiovascular systems are vulnerable to ultrafine particles. Investigation of ultrafine exposure for different types of vehicle and bicycle infrastructure is needed to understand how to lower exposures for commuters and protect public health.

Personal exposure studies have shown significantly increased ultrafine particle exposure concentrations associated with increased proximity to traffic and volume of traffic *(15-19)*. Traditionally, bicycle lanes have been placed adjacent to motor vehicle lanes. Recent designs in the U.S. have exchanged the locations of parallel parking and bicycle lanes- creating a "cycle track" - in which the cyclist is separated by a barrier (the parked cars) from the traffic stream. The barrier formed by the parked cars has the potential to create a perceptibly safer environment, reducing vehicle-bicycle collisions and attracting new riders who may otherwise feel unsafe biking next to moving vehicles. However, the full safety impact of cycle tracks (especially at intersections (*20*)) has not yet been empirically determined as they are a relatively new facility type (particularly in the U.S.). While the potential to reduce bicycle-vehicle conflicts has been the primary cited benefit of creating a cycle track, this study seeks only to determine if cycle tracks also can serve to lower ultrafine particle exposure concentrations. Results from the simultaneous assessment of traffic parameters and UFP exposure concentrations for a conventional bicycle lane and a cycle track are presented here.

METHODS

Measurements for this study were conducted on SW Broadway, a multi-lane, one-way southbound street in the downtown Portland core near the Portland State University campus. The road is used by bicyclists, cars, trucks, and buses. Traffic composition and volumes vary at this location throughout the day. Note that there is only one 4-leg intersection on this cycle track, all others are 3-leg since SW Broadway is adjacent to campus.

Prior to implementation of a cycle track design, the cross section consisted of three lanes with a traditional bicycle lane located between the right-most travel lane and a row of curb parking (see Figure 1(a)). After cycle track installation, two travel lanes remained, with an offset row of parallel parking providing a buffer to the cycle track, approximately 10 -11 feet in width (see Figure 1(b)). The arrow in Figure 1(b) points to the cycle track. The curb-to-curb distance was maintained during reconfiguration, requiring only lane re-striping, appropriate pavement markings, and new signage.

View of SW Broadway before cycle track View of SW Broadway with cycle track

(a) (b)

FIGURE 1 Cross-sectional configuration of SW Broadway (a) Prior to cycle track and (b) with cycle track implementation

After implementation of the cycle track, monitoring equipment was set up at a mid-block location, north of the intersection with SW Harrison Street (Figure 2). Particle number concentrations and traffic measurements were made over four days in the span of several months with different combinations of equipment and study durations depending on availability of equipment and personnel. On each study day, two P-trak ultrafine particle counters (TSI Model 8525) were placed in a parked car in the parallel parking (buffer) zone on the west side adjacent to the cycle track. P-trak instruments are commonly used in personal exposure studies of ultrafine particle for cyclists and other transportation modes because of portability and technological advances to measure number concentrations *(17)*. Number concentrations in ambient air are dominated by ultrafine particles. In comparable studies and personal exposure studies using the P-trak instrument, particle number concentrations and ultrafine particles are used interchangeably. Prior to data collection, a run of the P-trak instruments (recently factory calibrated) side-by-side in the laboratory for three and a half hours ensure instruments correlated (r^2= 0.996).

The parked car was utilized in a novel method to compare simultaneous measurements of exposure concentrations that would be experienced in a conventional bicycle lane versus a cycle track lane. The sensors were placed on the front seats of the car with the collection tube running out the windows, taped to the side-view mirrors (Figure 3). Measuring exposure on the driver's side of a car parked within this offset parking lane is representative of the exposure concentration in a traditional bicycle lane; exposure measured on the passenger-side represents the cycle track exposure concentration. The driver's side measurements are in the location and proximity to traffic where a bicycle lane would typically be marked and will be referred to as the bicycle lane results. The passenger-side measurements are located a few feet from the cycle track due to the white striped buffer area. The passenger side measurements are the upward limit for cycle track exposure concentrations due to the passenger-side-view mirror location and width of the cycle track. The cycle track UFP concentrations would range lower towards the sidewalk. Exposure concentration is a typical variable used in personal exposure studies to understand potential health impacts of humans in urban transportation microenvironments (*17*). Total or in-traffic exposure is the product of exposure concentration, exposure duration, and breathing rate.

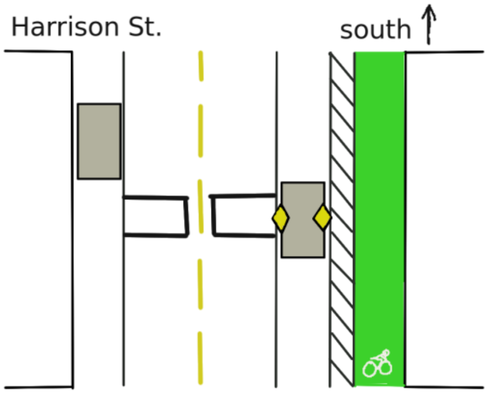

FIGURE 2 Study setup diagram. Green lane represents cycle track. Gray boxes represent cars. Yellow diamonds represent P-Trak instruments. Black lines in traffic lanes represent traffic counters.

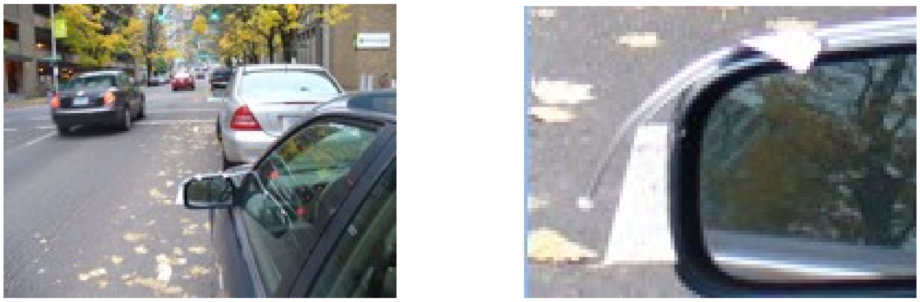
(a) **(b)**

FIGURE 3 Images of collection tube set-up on study vehicle (a) Driver's side-view mirror and one lane of moving traffic (b) Close-up of driver's side-view mirror and collection tube. Same setup used on passenger's side (not pictured).

All ultrafine particle counts were made at one-second resolution. The P-trak instrument measures particle number concentrations using condensation with isopropyl alcohol and an optical sensor. Particle number concentrations are obtained for particles in the size range of 0.02-1 µm. The particle counts measured by this instrument are dominated by the ultrafine particle size range. The maximum concentration level measured by this equipment is 500,000pt/cc.

Four different experimental setups were conducted, each described according to the study date and time periods in the following paragraphs. The first study design with P-Traks only was implemented on Nov. 24, 2009. Measurements at the first location began at 5:45AM and continued until 10:45AM. Particle exposure concentrations were measured in a second parking space from 10:58AM-1:52PM and in a third parking space from 2:05-4:51PM. Blocks in the City of Portland tend to be shorter than in most US cities. In all cases, the distance between P-track locations along Broadway did not exceed 50 feet.

Data collection on Feb. 8, 2010 occurred in the same parking space at the mid-block location from 5:31-10:49AM. Traffic data were also collected during this time period using MetroCount 5600 traffic tubes counters. The traffic counting tubes were placed in the right-most lane next to the vehicle containing the P-traks and collected individual vehicle records consisting of passage time, vehicle classification (based on length estimates), and speed.

Data collection on June 7, 2010 occurred in the same mid-block location as the first parking space on Nov. 24 and the Feb. 8 study day. Particle measurements occurred from 6:53AM-2:20PM. Additionally, a third P-trak was placed on the edge of the sidewalk closest to the cycle track in the same transect as the car P-traks from 7:54AM-2:20PM. Traffic tubes were placed across both lanes beginning at 5AM and traffic data were collected throughout the entire particle measurement period. The heights of the P-trak inlet tubes were maintained at the same elevation across the entire study period.

The final day of data collection occurred on July 13, 2010 from 7:25AM to 9:42PM. Particle measurements were made on the driver and passenger-sides of the study vehicle in the mid-block location. In this setup, traffic data were collected with traffic tube counters across both travel lanes.

RESULTS

Exposure Concentrations

Table 1 contains median and mean concentration values and ranges of exposure concentrations for the driver's side (traditional bicycle lane) and passenger's side (cycle track lane) positions for all study days.

One-sided paired t-tests were used to evaluate if the driver-side exposure concentrations were greater than the passenger-side exposure concentrations. T-test results and percent differences are shown in Table 1. Using a significance level of a p-value=0.05, exposure concentrations were significantly greater on the driver-side representing the typical bicycle lane compared to the passenger-side representing the cycle track facility for all study days.

While the bicycle lane exposure concentrations were always significantly greater than the cycle track exposure levels, there was a wide range in the mean of the differences and percent differences (8%-38%), see Table 1. The greatest difference (38%) between the bicycle lane and cycle track occurred for the second parking space from 10:58AM-1:52PM on Nov. 24. The next greatest difference (35%) occurred on the same day in the third space from 2:05-4:51PM. The time periods with greatest percent differences between the two bicycle facility designs overlap with time periods of high traffic volumes for SW Broadway. The smallest difference (8%) occurred on Feb. 8, 2010 from 5:31-10:49AM. The low volume of traffic in the first hour and a half of this study period would lead to less total ultrafine particle emissions and hence the smallest difference for the bicycle lane and cycle track measurements.

Particle number distributions showed bicycle lane measurements greater than 300,000-500,000pt/cc occurred more frequently compared to cycle track measurements. The inability of the equipment to capture peaks greater than 500,000pt/cc may have caused mean differences to be underestimated. These data suggests less peak exposure concentrations occur on the cycle track compared to a conventional bicycle lane since the cycle track measurements are the upper limit (due to cross-sectional location).

Not included in Table 1 are the results for the sidewalk measurements on June 7. The sidewalk median exposure concentration was equal to 12,900pt/cc with a mean concentration of 15,535pt/cc and a range from 6,890-433,000pt/cc. The bicycle lane concentrations were significantly greater than the sidewalk with a mean difference equal to 6,805pt/cc, t-value=28.4, p-value<0.01. The percent difference was 38%. The cycle track concentrations were also significantly greater than the sidewalk with a mean difference equal to 2,157pt/cc, t-value=20.5, p-value<0.01. The percent difference for the cycle track and sidewalk was 25%.

TABLE 1 Mean Number Concentrations, Ranges, Percent Differences, and t-test Results for Bicycle Lane and Cycle Track Exposure Concentration Comparisons

		Bicycle Lane			**Cycle Track**						
Date	**Time**	**Median (pt/cc)**	**Mean Conc (pt/cc)**	**Range (pt/cc)**	**Median (pt/cc)**	**Mean Conc (pt/cc)**	**Range (pt/cc)**	**Mean Diff. (pt/cc)**	**t-value**	**p-value**	**% Diff**
Nov24, 2009	5:45-10:45 AM	31,400	43,788	14,500-500,000	30,500	37,498	15,000-365,000	6,125	19.6	<0.01	15
Nov24, 2009	10:58 AM - 1:52 PM	28,200	56,845	4,510-500,000	26,000	35,802	13,600-500,000	21,043	28.8	<0.01	38
Nov24, 2009	2:05 - 4:51 PM	25,400	37,476	9,980-500,000	20,600	24,618	2,230-312,000	12,589	29.2	<0.01	35
Feb 8, 2010	5:31 -10:49AM	30,600	47,601	12,300-500,000	29,500	44,245	3,340-500,000	3,309	10.3	<0.01	8
June 7, 2010	6:53 AM -2:20 PM	14,700	25,271	3,340-500,00	14,200	20,805	5,750-500,000	4,465	20.9	<0.01	18
July 13, 2010	7:24 AM -9:42 PM	8,290	13,839	2,390-500,000	7,660	10,558	5,620-500,000	3,309	10.3	<0.01	24

Comparison with Measured Traffic

Traffic data were collected for 5 hours and 20 minutes from 5:31AM to 10:49AM on Feb. 8 during particulate matter collections. Traffic volume for the right-most travel lane during this period was 1,086 vehicles or 204 veh/hr/lane. Speeds for vehicles in this lane ranged from 6.40mph to 54mph with a time mean value of 30.11mph (Figure 5). Traffic composition was not analyzed in this paper.

Traffic increased throughout the morning peak period (with a maximum near 8:30AM), then remained relatively constant throughout the remaining time (Figure 4(a)). The steeper increase in traffic flow up until 8:15AM, followed by stabilization of the mean and greater variability in traffic flow may be due to the intersection reaching capacity or a change in intersection signalization timing as the morning progressed. Ultrafine particle number concentrations from the driver's side P-trak averaged at 5 minute intervals also show an increase up to a peak in a Loess smoothing curve around approximately 8:15AM (Figure 4(b)). Exposure concentration differences between the bicycle lane and cycle track show a peak around 8:40-8:45AM (Figure 4(c)).

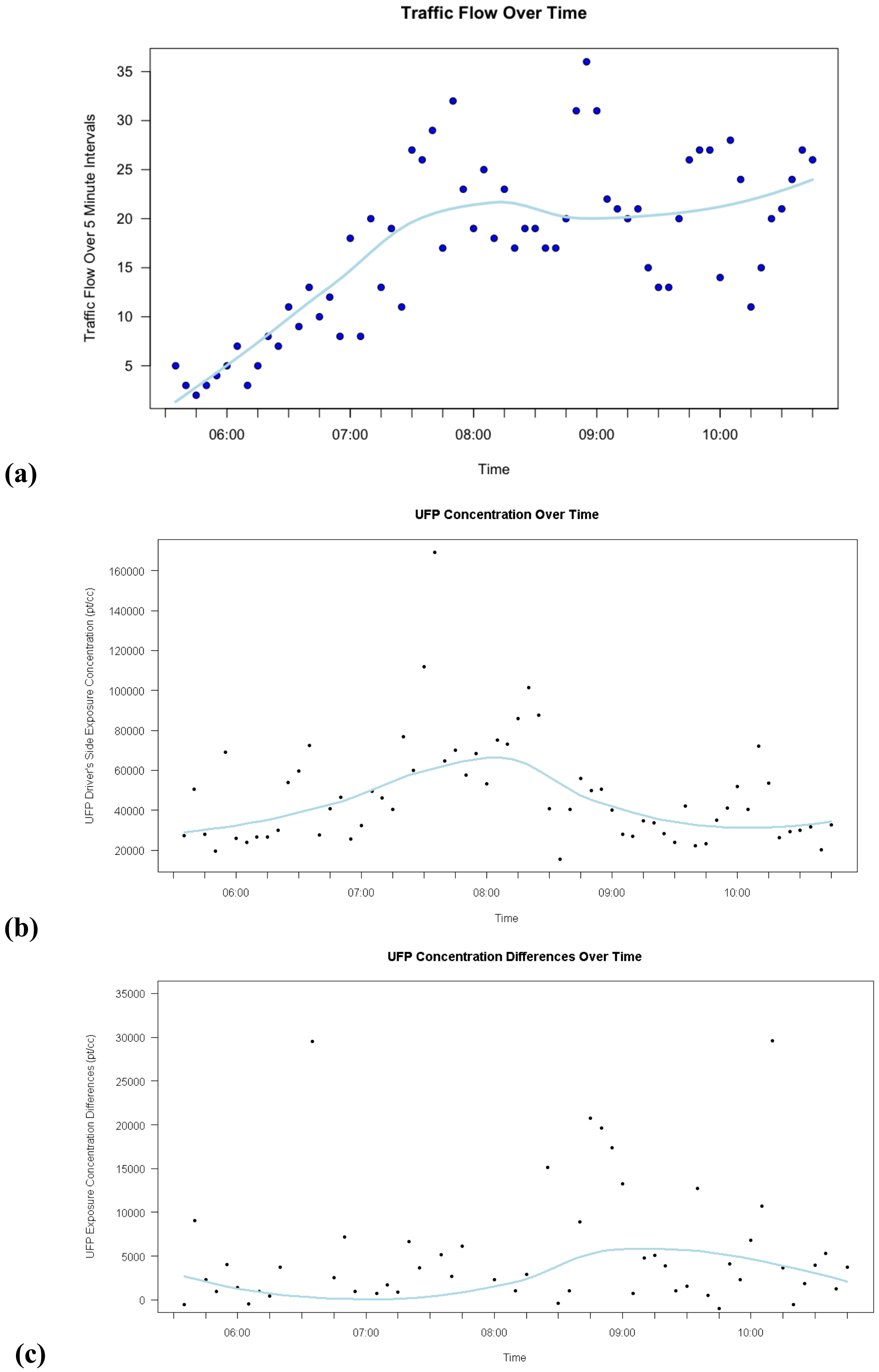

FIGURE 4 Feb. 8 (a) Traffic flow per 5-mim intervals versus time (b) UFP concentrations from driver's side averaged over 5-minute intervals versus time (c) UFP concentration differences between bicycle lane and cycle track sides averaged over 5-min intervals versus time. All lines represent Loess smoothing curves.

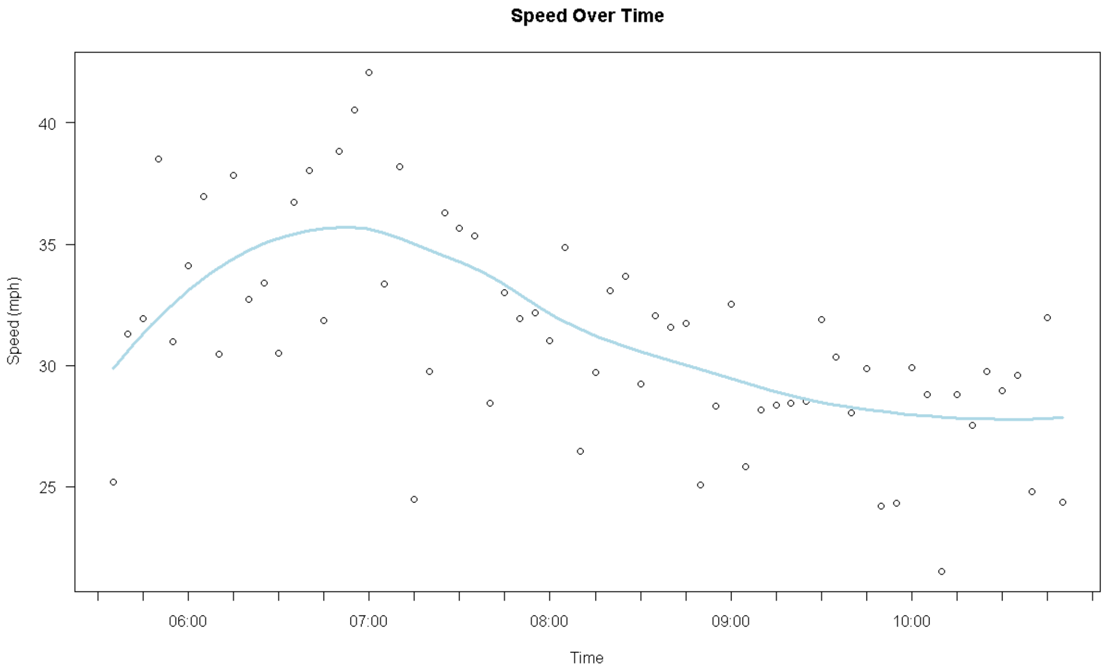

FIGURE 5 Feb. 8 Speed averaged over 5min intervals with a Loess smoothing curve.

Traffic data obtained on June 7 were invalid due to a data collection error. Traffic data for July 13 were collected for approximately 14 hours, including the morning and evening periods. The total traffic count from 7:25AM to 9:42PM across both lanes was 8,232 vehicles or 294veh/hr/ln.

Traffic increased relatively linearly from 10:15AM until a peak around 4:15PM as shown by a Loess smoothing curve in Figure 6(a). Traffic declined through the rest of the evening until the tubes were disconnected. Ultrafine particle concentrations from the driver's side averaged over a 5 minute interval show an increase up to a point around noon (Figure 6(b)). Figure 6b shows the variability or range of the ultrafine particle exposure concentrations around the Loess curve to be greater during the early and middle parts of the day compared to the end of the day when traffic volumes were decreasing. Exposure concentration differences also show a peak at noon (Figure 6(c)).

On July 13th, the time mean speed of vehicles in the right-most motor vehicle travel lane (adjacent to research vehicle) was 28.34 mph, with a range from 1.20 mph to 53 mph. The left-most travel lane (furthest from the cycle track and study vehicle) had a time mean speed of 25.83 mph with a range from 5.70 to 56.50 mph. Both lanes together averaged 27.62 mph, with a range from 11 to 44.80 mph.

The averaged speeds over five minute intervals of vehicles in both lanes did not fluctuate much through the day with the Loess smoothing curve not deviating far from the range of 25mph to 32mph (Figure 7). The decreasing trend in speed in the morning from 7:30-11AM seen on Feb. 8 was also seen on July 13 (Figure 5 and 7). This trend continued on July 13 until the median speed dipped to about 25mph from 1:30-2:30PM. Speed began to increase linearly at about 5PM on July 13. Traffic counts peaked around 4:15PM, so the time periods with fewer cars on the road followed the slight increase in car speeds.

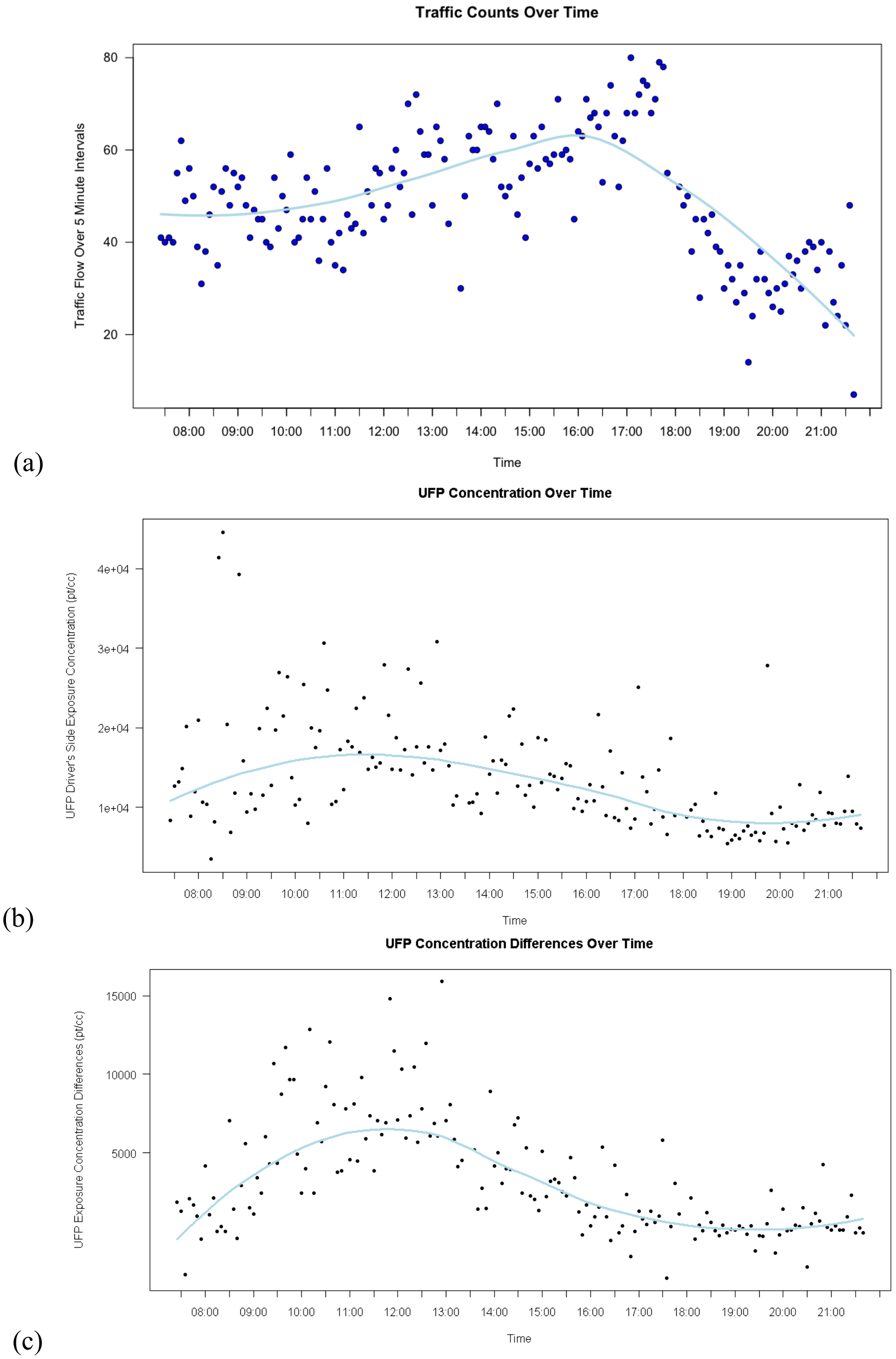

FIGURE 6 July 13 (a) Traffic flow per 5-mim intervals versus time (b) UFP concentrations from driver's side averaged over 5-minute intervals versus time (c) UFP concentration differences between bicycle lane and cycle track sides averaged over 5-min intervals versus time. All lines represent Loess smoothing curves.

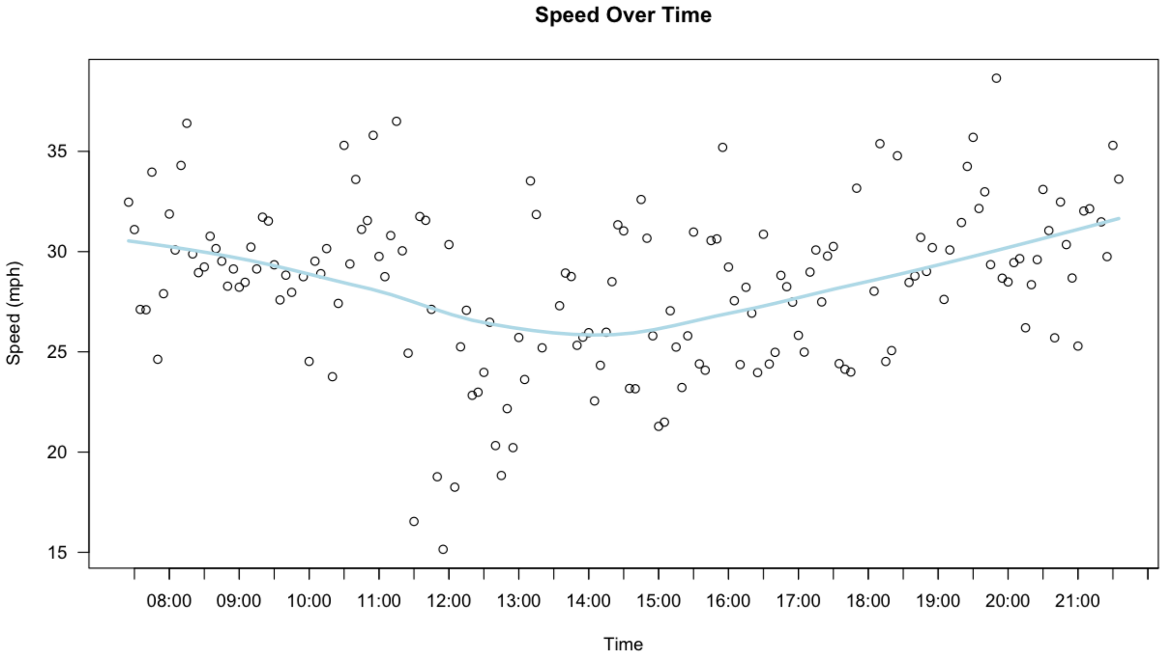

FIGURE 7 July 13 Speed averaged over 5-minute intervals versus time with a Loess smoothing curve.

Analysis of the individual traffic variables to UFP levels using regression and functional optimization techniques did not result in a statistically significant relationship. The results of this analysis suggest that the interaction of traffic speed and traffic counts alone cannot functionally account for the data measured in this study. Traffic composition and wind measurements are also likely needed to understand the functional relationship between traffic and UFP levels at this study site and are to be investigated in further studies.

DISCUSSION

Ultrafine particle exposure concentrations were significantly greater on the driver's side than the passenger's side for all study days. The one-second sampling interval captures very quick changes and short term peak exposures explaining the wide range of particle number concentrations for the bicycle lane and cycle track positions. The cycle track has the potential to lower ultrafine exposure concentrations compared to a traditional bicycle lane.

The differences in the ultrafine particle levels for the typical bicycle lane and cycle track are most likely due to the increased horizontal distance from the traffic stream and the airflow over the parked vehicle. Over this distance ultrafine particles coagulate *(21)* and grow to larger, potentially less harmful particles. It is unlikely that the parked cars act as a physical barrier for the ultrafine particles to which particles collide with the car surfaces and adhere to them. Ultrafine particles behave as a gas and this explanation would relate more appropriately to larger particles with greater mass. However, future studies will test dry deposition of ultrafine particles for the possibility of additional explanation. The possibility of a traffic-pollution "shadow" on the passenger-side of the car where the cycle track collection tube intake was located will be evaluated in future work using a computational fluid dynamic model to generate wind fields

The continued significant decline in exposure concentrations from bicycle lane to cycle track to sidewalk also shows a strong likelihood of horizontal distance being the mechanism for the exposure level differences. An assessment of pedestrian exposure to air pollutants along a major road in central London, UK, found ultrafine particle number counts to be significantly higher when walking along the curb side edge of the sidewalk compared to the building side (22). The width of the sidewalk is comparable to the width of the parking lane and buffer zone placed between the cycling lane and motor vehicles in the cycle track design.

The placement of the study vehicle from 10:58AM to 1:52PM on Nov. 24 was different than the mid-block location just north of SW Harrison used on all other study days. For this time period, the vehicle was at the front parking spot closest to the traffic light at the intersection north of SW Harrison. This time period showed the greatest mean and percent difference for the bicycle lane and cycle track concentrations. Future studies should further investigate the effect of proximity to signalized intersections and signal queuing on ultrafine particle concentrations. Placing study vehicles in differing proximities to intersections, along with enhanced traffic monitoring, may lead to a better understanding of geometric and traffic effects on ultrafine particle exposures.

Traffic data from Feb. 8 and July 13 indicate a traffic pattern on SW Broadway of increasing traffic beginning at 5:30AM, elevated traffic flows past the morning peak period into the afternoon (10:45AM-4:00PM), and a decline in traffic flows beginning at 5:00PM (Figure 4(a) and 6(a)). The greatest exposure concentration differences of 38% and 35% (Table 1) for the two bicycle facilities occurred during 10:45AM-1:52PM and 2:05-4:51PM within the time period of elevated traffic flows. The highest exposure concentration differences from Figure 4(c) and Figure 6(c) occur around 8:45AM and 12:00PM also within the elevated traffic flow pattern. Figure 6(c) shows decreased exposure concentration differences from 7:00-8:00PM during a time period of declining traffic and lowest traffic flows. These results begin to indicate the greatest exposure level differences for the bike facilities occur when traffic was greatest. Future work will continue to collect full-day traffic and air quality measurements to track this relationship of higher exposure concentration differences associated with higher traffic levels.

A count of bicyclists prior to installation of the cycle track found that bicycle volumes peaked around 9:00AM and again at 5:30PM (around 60 bicycles per hour). The time spans of elevated motor vehicle traffic and bicyclist traffic overlap on SW Broadway. The above results suggest that cycle track facilities have the greatest potential to mitigate ultrafine particle exposures for bicyclists on roadways and transportation environments with concurrently high auto use and cyclist activity.

The traffic flow peak around 4:00PM on July 13 was not matched by a peak in UFP, which were declining from a peak around mid-day (Figure 6(a) and 6(b)) suggesting the data may be missing an important correlate such as wind parameters. Future work with radar and video to capture traffic composition and the use of 3-dimensional ultrasonic anemometers that measures vertical and horizontal wind fluxes will allow for further exploration into such effects.

CONCLUSION

An original method was developed to measure and compare simultaneous ultrafine particulate exposure for cyclists in a traditional bicycle lane and a cycle track. Ultrafine particle number concentrations were significantly higher in the typical bicycle lane than the cycle track for all study days, and nearly all study periods within those days. The higher frequency of exposure concentrations greater than 300,000-500,000pt/cc in the bicycle lane compared to the cycle track suggests a cyclist may encounter fewer peak exposure concentrations in the cycle track. Additionally, the cycle track measurements in this study are the upper limit due to cross-sectional location. Significantly lower ultrafine number concentrations measured on the cycle track are attributable to the increased distance from the motorized traffic provided by the cycle track configuration. Increasing the bicycle facility distance from traffic sources is difficult in cities with set road widths. A cycle track with a parking lane buffer offers a realistic solution for roads in urban areas with parking lanes to potentially lower ultrafine exposures for cyclists.

Traffic measurements showed the exposure concentration differences to be greatest at times of highest traffic volumes, emphasizing the importance of mitigation techniques in areas with simultaneously high volumes of motor vehicle and bicycle commuters. Initial findings show possible effects of proximity to signalized intersections on increased ultrafine particle exposure concentration differences for a bicycle lane and cycle track. These elements need to be studied in further detail along with local wind and more temporal and seasonal measurements of traffic and associated ultrafine particle exposure levels.

The findings of this study show a cycle track roadway design may be more protective for cyclists than a traditional bicycle lane in terms of lowering exposure concentrations of ultrafine particles. This, of course, must be balanced against other consideration such as vehicle-bicycle conflicts at intersections and other design considerations. Based on these initial findings, understanding roadway and traffic effects on exposure levels can help guide bicycle facility design and pinpoint locations in which mitigation of exposure levels by placement of facilities such as cycle tracks may be most important.

ACKNOWLEDGEMENTS

The authors acknowledge the Miller Grant Foundation and the Oregon Transportation Research and Education Consortium (OTREC) for funding this work.

REFERENCES

1. Zhang, J. and L. Morawska. Combustion sources of particles: 2. Emission factors and measurement methods. *Chemosphere*, 49, 2002, pp. 1059-1074.
2. Kittelson, D.B. Engines and nanoparticles: a review. *Journal of Aerosol Science*, 29, 1998, pp. 575-588.
3. Ristovsk, Z.D., Morawska, L., Bofinger, N.D., and J. Hitchins. Submicrometer and Supermicrometer Particles from Diesel Vehicle Emissions. *Environmental Science and Technology*, 32, 1998, pp. 2033-2042.
4. Harrison, R.M., Jones, M., and G. Collins. Measurements of the physical properties of particles in the urban atmosphere. *Atmospheric Environment*, 33, 1999, pp. 309-321.
5. Junker, M., Kasper, M., Roosli, M., Camenzind, M., Kunzli, N., and C. Monn. Airborne particle number profiles, particle mass distributions and particle-bound PAH

concentrations within the city environment of Basel: an assessment part of the BRISKA Project. *Atmospheric Environment*, 34, 2000, pp. 3171-3181.

6. McNabola, A., Broderick, B.M., and L.W. Gill. Optimal cycling and walking speed for minimum absorption of traffic emissions in the lungs. *Journal of Environmental Science and Health, Part A, Toxic/hazardous substances & environmental engineering*, 42, 2007, pp. 1999-2007.
7. Van Wijnen, J.H., Verhoeff, A.P., Jans, H.W.A., and M. van Bruggen. The exposure of cyclists, car drivers, and pedestrians to traffic-related air pollutants. *International Archives of Occupational and Environmental Health*, 67, 1995, pp. 187-193.
8. O'Donoghue, R.T., Gill, L.W., McKevitt, R.T., and B.M. Broderick. Exposure to hydrocarbon concentrations while commuting or exercising in Dublin. *Environmental International*, 33, 2007, pp. 1-8.
9. McNabola, A., Broderick, B.M., and L.W. Gill. Relative exposure to fine particulate matter and VOCs between transport microenvironments in Dublin: Personal exposure and uptake. *Atmospheric Environment*, 42, 2008, pp. 6496-6512.
10. Harrison, R.M., Shi, J.P., Xi, S. Khan, A., Mark, D., Kinnersley, R. and J. Yin. Measurement of number, mass, and size distribution of particles in the atmosphere. *Philosophical Transactions: Mathematical, Physical, and Engineering Sciences*, 358, 2000, pp. 2567-2580.
11. Møller, P., Folkmann, J. K., Forchhammer, L., Bräuner, E.V. Danielsen, P.H., Risom, L. and S. Loft. Air pollution, oxidative damage to DNA, and carcinogenesis. *Cancer Letters*, 266, 2008, pp. 84-97.
12. Vinzents, P.S. Moller, P., Sorensen, M., Knudsen, L.E., Hertel, O., Jensen, F.P., Schibye, B., and S. Loft. Personal Exposure to Ultrafine Particles and Oxidative DNA Damage. *Environmental Health Persepctives*, 113, 2005, pp. 1485-1490.
13. Seaton, A., Godden, D., MacNee, W. and K. Donaldson. Particulate air pollution and acute health effects. *The Lancet*, 345, 1995, pp. 176-178.
14. Li, N., Sioutas, C., Cho, A., Schmitz, D., Misra, C., Sempf, J., Wang, M., Oberley, T., Froines, J., and A. Nel. Ultrafine particulate pollutants induce oxidative stress and mitochondrial damage. *Environmental Health Perspectives*, 111, 2003, pp. 455-460.
15. Thai, A., McKendry, I. and M. Brauer. Particulate matter exposure along designated bicycle routes in Vancouver, British Columbia. *Science of the Total Environment*, 405, 2008, pp. 26-35.
16. Boogaard, H., Borgman, F., Kamminga, J., and G. Hoek. Exposure to ultrafine and fine particles and noise during cycling and driving in 11 Dutch cities. *Atmospheric Environment*, 43, 2009, pp. 4234-4242.
17. Kaur, S., Nieuwenhuijsen, and R.N. Colvile. Fine particulate matter and carbon monoxide exposure concentrations in urban street transport microenvironments. *Atmospheric Environment*, 41, 2007, pp. 4781-4810.

18. Berghmans, P., Bleux, N., Int Panis, L., Mishra, V.K., Torfs, R., and M. Van Poppel. Exposure assessment of a cyclist to PM_{10} and ultrafine particles. *Science of the Total Environment*, 407, 2009, pp. 1286-1298.
19. Kaur, S. and M.J. Nieuwenhuijsen. Determinants of Personal Exposure to $PM_{2.5}$, Ultrafine Particle Counts, and CO in a Transport Microenvironment. *Environmental Science and Technology*, 43, 2009, pp. 4737-4743.
20. Jensen, S.U., Rosenkilde, C. and N. Jensen. Road safety and perceived risk of cycle facilities in Copenhagen. Produced for the Municipality of Copenhagen. Produced by Trafitec. http://www.ecf.com/files/2/12/16/070503_Cycle_Tracks_Copenhagen.pdf
21. Yao, X., Lau, N.T., Fang, M., and C.K. Chan. Real-Time Observation of the Transformation of Ultrafine Atmospheric Particle Modes. *Aerosol Science and Technology*, 39, 2005, pp. 831-841.
22. Kaur, S., Nieuwenhuijsen, M.J., and R.N. Colvile. Pedestrian exposure to air pollution along a major road in Central London, UK. *Atmospheric Environment*, 39, 2005, pp. 7307-7320.